

Impacts of long-term nitrogen addition on nitrous oxide in a temperate grassland

Si Chen^{1,4}, Tianpeng Gao^{2,3}, Tianxiang Hao⁴, Kaihui Li⁵, Xuejun Liu^{2*}

¹ Lanzhou City University, Lanzhou 730070, China.

²School of Biological and Environmental Engineering, Xi'an University, Xi'an 710065, China

³Engineering Center for Pollution Control and Ecological Restoration in Mining of Gansu Province, Lanzhou City University, Lanzhou 730070, China

⁴ Key Laboratory of Plant-Soil Interactions of MOE and Beijing Key Laboratory of Farmland Pollution Prevention and Remediation, College of Resources and Environmental Sciences, China Agricultural University, Beijing 100193, China.

⁵ Key Laboratory of Biogeography and Bioresource in Arid Land, Xinjiang Institute of Ecology and Geography, Chinese Academy of Sciences, Urumqi 830011, China

Abstract: Atmospheric nitrogen (N) deposition has increased dramatically due to increased human activities since the industrial revolution. However, it is still unclear what the responses of soil nitrous oxide (N₂O) is to long-term elevated N deposition in a temperate grassland. Here, we conducted an *in situ* field experiment to investigate these responses to long-term high N addition on a temperate steppe in Inner Mongolia, China, from April 2017 to October 2018. Soil N₂O emissions significantly increased by long-term N addition, use of structural equation modeling (SEM) showed that topsoil (0-5 cm) NH₄⁺-N content was the most important limiting factor for N₂O emission. Our results indicate that long-term high N addition showed a significantly increase in N₂O emission in this temperate grassland.

1 Introduction

Anthropogenic nitrogen (N) enrichment and global warming are primary components of global change. N deposition to terrestrial ecosystems has increased from 17.4 Tg N yr⁻¹ in late 1860s to 187 Tg N yr⁻¹ in the 2005, and is expected to increase to 200 Tg N·yr⁻¹ in 2050^[1]. This in natural ecosystems has many negative effects on ecosystem services: biodiversity loss, nutrient imbalance, and increased greenhouse gases (GHGs) emissions^[2,3]

In Inner Mongolia grassland occupies a large area of northern China and 13% of the global grassland area^[4]. However, models of N₂O emission with the N deposition shows N₂O emission increased as exponential and unimodal response^[5,6], when the N reached biological threshold, N₂O emission transitions to the N saturation stage, at this stage, higher N could not increase N₂O emission, including in some temperate and subtropical grasslands^[7].

N saturation model believes that in natural ecosystems, enhanced N deposition would produce more soil N₂O^[8], but recent study on steppe showed that influence of significant climatic factors on N₂O depends on the ability to alleviate N₂O emission limited by soil "nitrogen deficiency"^[9,10]. Therefore, the effects of long-term N addition on N₂O emission are still uncertain. In another

way, N enrichment was always accompanied by the changes in soil properties like soil acidification, available N and C states, as well as vegetation of ecosystem. Therefore, it is also valuable to find the complex response of N₂O fluxes of soil to variations in multiple indirect factors caused by N deposition of grassland^[11].

Therefore, we set up a long-term *in situ* experiment in Inner Mongolia, China in 2005, which has caused severe soil acidification after 13 years of N fertilization, and significant changes in soil properties^[12]. To assess how these changes have affected soil N₂O emission, we measured N₂O fluxes with high frequency from May 2017 to October 2018, with the objectives of: (i) quantifying the response of the N₂O emission to long-term increased N addition, and (ii) identifying the main controlling factors of N₂O caused by long-term increased N addition.

2 Experimental methods

2.1 Site description

This study was conducted in a semi-arid temperate grassland in Duolun County (116°17'E, 42°02'N, elevation 1324 m), northeastern of Xilingol of Inner Mongolia, China. The rainfall and air temperature average 316 mm and 3.3°C between 2006 to 2016, and 81% of total

* Corresponding author: Xuejun Liu, Email: liu310@cau.edu.cn

precipitation in growing season (May 10th to August 30th; Fig.1(a)). The local N deposition has increased to 15.4 kg N ha⁻¹ yr⁻¹ in 2015, consist with 58% as NH₄⁺ and 42% as NO₃⁻. 0-10 cm layer soil pH and bulk density are 7.12 and 1.31 g cm⁻¹. Soil total N and total P contents are 12.3 g kg⁻¹ and 1.3 g kg⁻¹, organic carbon and C:N are 0.28 g kg⁻¹ and 21, respectively.

2.2 Experimental treatments

The experiment was designed with three N addition treatments of four replicates that were arranged randomly. Plot area was 5×5 m with a 1 m wide buffer zone. Ammonium nitrate was applied as foliar fertilizer at simulated N deposition rates of N₀ (0 kg N ha⁻¹ yr⁻¹), N₃₀ (30 kg N ha⁻¹ yr⁻¹) and N₆₀ (60 kg N ha⁻¹ yr⁻¹). The fertilizer was divided into 2 equal portions, and sprayed evenly in mid-June and mid-July every year (15th June and 17th July in 2017). N₂O was measured from May 2017 to October 2018.

2.3 Measurements method

Static closed chamber technique was used to measure N₂O fluxes. The chambers were made of removable top and previously-installed stainless steel (into soil for 10 cm). Sampling was conducted between 10:00 am and 12:00 pm. Gases fluxes were collected once or twice a week. Gas samples were injected into 50 ml, and were analyzed using a gas chromatograph (GC; Agilent 7890A, Agilent Technologies, Santa Clara, CA, USA) equipped with an electron capture detector for determination of N₂O^[13,14].

Fresh soil samples were extracted in 0.01 mol L⁻¹ CaCl₂ solution and shaken for 1 hour. The suspension was analyzed for nitrate and ammonium concentration using a continuous-flow auto-analyzer (Seal AA3, Germany). Dissolved organic carbon (DOC) was extracted from the fresh soil sub-samples by shaking with deionized water, and filtered at 0.45 μm and the filtrate analyzed with a TOC analyzer (multi N/C 3100, Jena, Germany). Soil pH was measured using a pH Meter (Seven Easy, Mettler-Toledo, Switzerland). Soil water content and temperature in the 0-5 cm layer were measured using a TDR probe.

3 Statistical analyses

Treatment enrichment effects on soil pH, NO₃⁻-N, NH₄⁺-N content, DOC, total emission of N₂O were assessed using a least significant difference test (LSD, *P* < 0.05). In addition, the relationships between fluxes of N₂O with soil moisture, soil NO₃⁻-N and NH₄⁺-N contents, pH were analyzed using structural equation models (SEM). One-way repeated measures analysis of variance was used to analysis. All statistical analyses were conducted using SPSS (version 20.0) with statistically significant differences set at *p* < 0.05. SEM analyses were carried out using AMOS 22.0. All figures were drawn using Sigmaplot (version 13.0).

4 Results

Long-term high N addition increased significantly soil NO₃⁻-N, DOC content (except in N₃₀ Treatment), and decreased pH significantly by long-term N addition, but has no effect on NH₄⁺ content.

Small N₂O emissions were observed from N₀ plots, ranging from 0.59 to 16.32 μg N m⁻² h⁻¹ (average annual 4.45 ± 0.34 μg N m⁻² h⁻¹) (Fig. 1b). Long-term increased N addition significantly increased the N₂O emissions (*P* < 0.001) (Fig 1b; Table 3), with average annual emissions from N₃₀ and N₆₀ plots of 9.49 ± 0.79 and 20.35 ± 3.23 μg N m⁻² h⁻¹, respectively (Fig 1b). The SEM showed that soil moisture had a significantly positive correlation with soil N₂O emissions both directly and indirectly, and that soil NH₄⁺-N content was the most important influencing factor (Fig. 2). Notably, neither soil pH or DOC had a significant impact on N₂O emissions (Fig. 2).

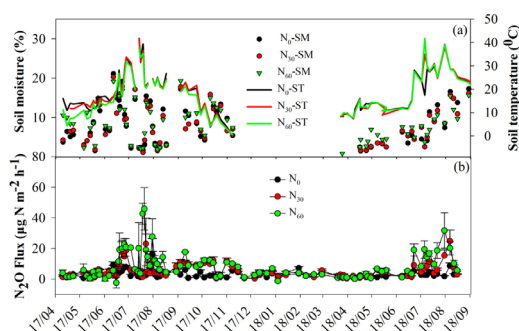


Figure 1. The Variation from April 2017 to October 2018 in soil moisture and soil temperature (°C) (a), and fluxes in N₂O (b) (mean, n=4) in an Inner Mongolian grassland.

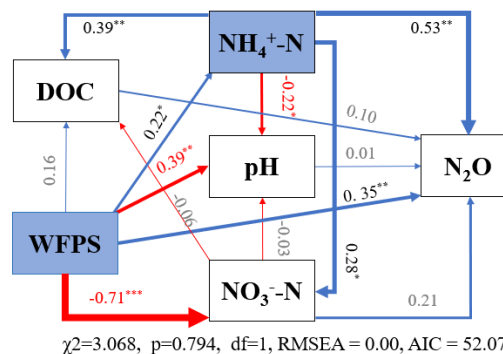


Figure 2. The effects on N₂O emission (n=28) of soil moisture, temperature, pH, soil NO₃⁻-N content and NH₄⁺-N content as tested by structural equation modeling (SEM). Single-headed arrows showed that the impact of different significant controls on N₂O emission was analyzed. The blue arrows represent positive effects, and red arrows negative effects. The width of the arrows indicated the strength of the relationship. The numbers are standardized path coefficients, which show the significance of the variables in the model. Goodness-of-fit statistics for the model are shown below the model. *, ** and *** indicate significant effect at *P* < 0.05, *P* < 0.01, and *P* < 0.001, respectively.

Table1. The variation in NO₃⁻-N, NH₄⁺-N, soil pH, dissolved organic carbon (DOC) of topsoil (0-5 cm) with the treatments in August 2017. Values in the same column with different letters indicate significant differences (P < 0.05)

Treatments	NH ₄ ⁺ -N (mg kg ⁻¹)	NO ₃ ⁻ -N (mg kg ⁻¹)	pH	DOC (mg kg ⁻¹)
N ₀	5.30±2.82a	11.65±5.28a	6.32±0.09b	68.46±4.01a
N ₃₀	5.97±0.95a	17.29±2.05a	5.91±0.07a	75.83±8.89a
N ₆₀	6.79±1.30a	19.82±3.90b	5.94±0.35a	82.59±8.06b

Table2. Repeated measures ANOVA analysis of variance with F and P values of the total effects of long-term nitrogen (N) addition on N₂O emissions in long-term N addition; n indicates sample size.

Source of variation	N ₂ O		
	df	F	P
Date	1.834	9.861	0.009**
N	1	101.127	<0.001**
Date*N	1.834	2.603	0.142

5 Discussion

N₂O emissions of the temperate grassland (Fig. 1b; Table 3), acted as a small source of N₂O and the cumulative emission of N₂O from N₀ plots is 0.37 kg N ha⁻¹ (Fig. 1b), higher than that from other ungrazed steppes in Inner Mongolia (below 0.28 kg N ha⁻¹)^[16]. N₂O emissions was triggered by the day following fertilization, and the peaks lasted about 20 days and the emissions fell to the baseline (Fig. 1b), which was related to the dynamics of soil mineral N (Table 1). The SEM showed that NH₄⁺-N was the most important controlling factor for N₂O emission rather than NO₃⁻-N or pH (Fig. 2), indicated that N₂O emission was mainly as byproduct of soil ammonia oxidation. Dry weather and the sandy texture (sand 62.8%) had been keeping soil WFPS below 30% (Fig. 1a) over the observed period, in that conditions N₂O by denitrification produced much less (WFPS≤70%)^[17]. Recent studies suggested that in a typical steppe, long-term nitrogen fertilization and soil pH could have contrasting effects on ammonia oxidizing microorganisms^[18], our result showed that pH had less influence to N₂O, that was primarily due to N-reduced pH did not dramatically shifted N inducing function and diversity of ammonia oxidizers of grassland^[19].

6 Conclusion

Atmospheric N deposition has dynamically increased in recent years, which will significantly affect the N₂O balance in grassland ecosystems. Long-term (13 year) increased N addition significantly increased N₂O emission. Soil NH₄⁺-N content and soil moisture were the most important limiting factors for N₂O emissions, pH had less influence to N₂O. We conclude that N₂O emission would be enhanced by long-term increased N deposition, which will have an important effect for climate change with elevating atmospheric N deposition in future.

Acknowledgments

This work was supported by the Doctoral Research Fund of Lanzhou City University (LZCU-BS2019-43), Natural Science Foundation of China (NSFC 31860176), Key Research and Development Program of Gansu Province (20YF3FA037), Key Research and Development Program of Shanxi Province (2020ZDLSF0606) and XAWLK YTD012 the Natural Science Foundation of Gansu Province, China (21JR1RA319). Chinese Key State Research and Development Programme (2017YFC0210100), China National Funds for Distinguished Young Scientists (Grant No. 41425007), and the UK Newton Fund through the BBSRC project ‘China Virtual Joint Centre for Improved N Agronomy (CINAG)’ (BB/N013468/1)

References

- Galloway, J. N., Townsend, A. R., Erisman, J. W., Bekunda, M., Cai, Z. C., Freney, J.R., Martinelli, L. A., Seitzinger, S. P., Sutton, M. A. (2008). Transformation of the nitrogen cycle: recent trends, questions, and potential solutions. *Science*, 320, 889-982.
- Aber, J., McDowell, W., Nadelhoffer, K., Magill, A., Berntson, G., Kamakea, M., McNulty, S., Currie, W., Rustad, L., Fernandez, I. (1998). Nitrogen saturation in temperate forest ecosystems - Hypotheses revisited. *Bioscience*, 48, 912-93.
- Lu, X. K., Mao, Q. G., Gilliam, F. S., Luo, Y. Q., Mo, J. M. (2015). Nitrogen deposition contributes to soil acidification in tropical ecosystems. *Global Change Biology*, 20, 3790-3801.
- Wang, Y. S., Xue, M., Zheng, X. H., Ji, B. M., Du, R., Wang, Y. F. (2005). Effects of environmental factors on N₂O emission from and CH₄ uptake by the typical grasslands in the Inner Mongolia, *Chemosphere*, 58(2), 205-215.
- Hao, T. X., Song, L., Goulding, K., Zhang, F. S. and Liu, X. J. (2018). Cumulative and partially recoverable impacts of nitrogen addition on a temperate steppe. *Ecology Application*, 28, 237-248.
- Zhang, W., Liu, C. Y., Zheng, X. H., Fu, Y. F., Hu, X. X., Cao, G. M., ButterbachBahl, K. (2014). The increasing distribution area of zokor mounds weaken greenhouse gas uptakes by alpine meadows in the Qinghai-Tibetan Plateau. *Soil Biology and Biochemistry*, 71, 105-112.

7. Smith, K. A., Ball, T., Conen, F., Dobbie, K., Massheder, J. R. (2003). Exchange of greenhouse gases between soil and atmosphere: interactions of soil physical factors and biological processes. *European Journal of Soil Science*, 69(1), 10-20.
8. Zhang, W., Mo, J. M., Yu, G. R., Fang, Y. T., Li, D. J., Lu, X. K., Wang, H. (2008). Emissions of nitrous oxide from three tropical forests in Southern China in response to simulated nitrogen deposition. *Plant and Soil*, 306(1-2), 221-236.
9. Aronson, E. L., Helliker, B. R. (2010). Methane flux in non-wetland soils in response to nitrogen addition: a meta-analysis. *Ecology*, 91, 3242-3251.
10. Yue, P. Li, K. H., Gong, Y. M., Hu, Y. K., Mohammat, A., Christie, P., Liu, X. J. (2016). A five-year study of the impact of nitrogen addition on methane uptake in alpine grassland. *Scientific Reports*, 6, 32064.
11. Tilman, D., Knops, J., Wedin, D., Reich, P., Ritchie, M., Siemann E. (1997). The influence of functional diversity and composition on ecosystem processes. *Science*, 277(5330), 1300-1302.
12. Zeng, J., Liu, X. J., Song, L., Lin, X. G., Zhang, H. Y., Shen, C. C., Chu, H. Y. (2016). Nitrogen fertilization directly affects soil bacterial diversity and indirectly affects bacterial community composition. *Soil Biology and Biochemistry*, 92, 41-49.
13. Xu, W., Zhao, Y.H., Liu, X. J., Dore, A. J., Zhang, L., Liu, L., Cheng M. M. (2018). Atmospheric nitrogen deposition in the Yangtze River basin: Spatial pattern and source attribution. *Environmental Pollution*, 232, 546-555.
14. Liu, C. Y., Zheng, X. H., Zhou, Z. X., Han, S. H., Wang, Y. H., Wang, K., Liang, W. G., Li, M., Chen, D. L., Yang, Z. P. (2010). Nitrous oxide and nitric oxide emissions from an irrigated cotton field in Northern China. *Plant and Soil*, 322, 123-134.
15. Jiang, X., Cao, L., Zhang, R. (2014). Changes of labile and recalcitrant carbon pools under nitrogen addition in a city lawn soil. *Journal of Soils Sediments*, 14, 515-524.
16. Shang Q. Y., Yang X. X., Gao C. M., Wu P. P., Liu, J. J., Xu, Y. C., Shen, Q. S., Zou, J. W., Guo, S. W. (2011). Net annual global warming potential and greenhouse gas intensity in Chinese double rice-cropping systems: a 3-year field measurement in longterm fertilizer experiments. *Global Change Biology*, 17(25), 2196-2210.
17. Zhu, X., Burger, M., Doane, T. A. and Horwath, W. R. (2013). Ammonia oxidation pathways and nitrifier denitrification are significant sources of N₂O and NO under low oxygen availability. *Proceedings of the National Academy of Sciences of United States of America*, 110(16), 6328-6333.
18. Ying, J. Y, Li, X. X., Wang, N. N., Lan, Z. C., He, J. Z., Bai, Y. F. (2017). Contrasting effects of nitrogen forms and soil pH on ammonia oxidizing microorganisms and their responses to long-term nitrogen fertilization in a typical steppe ecosystem. *Soil Biology and Biochemistry*. 107, 10-18]
19. Chen, S., Hao, T. X., Goulding K., Misselbrook, T., Liu, X. J. (2019). Impact of 13-years of nitrogen addition on nitrous oxide and methane fluxes and ecosystem respiration in a temperate grassland. *Environmental Pollution*, 252, 657-681. doi.org/10.1016/j.envpol.2019.03.069